

Lake Illawarra Estuary Health and Water Quality Report 2024

Reporting on data May 2023 to April 2024

Prepared by Wollongong City Council

September 2024







Summary

Wollongong City Council and Shellharbour City Council have been monitoring water quality and estuary health in Lake Illawarra since October 2013. This is a report for the monitoring period May 2023 to April 2024, focussing on those measurements that are especially important for assessing the lake condition for its ecosystem health and recreational use.

Rainfall and hence freshwater discharge continues to be an influential factor on the water quality of the Lake. The total rainfall for the sampling period was significantly lower than the last two years. In comparison to the last sampling period 2022-23 where the majority of the rain fell over the summer period the overwhelming majority of the rain fell over the winter period during 2023-24. The analysis of estuary health is undertaken with data from the summer period (1st November 2023- 30th April 2024) and nine of the eleven sites were rated as good or very good. The overall grade for the Lake Illawarra as a whole was calculated, with the Lake receiving a Good rating or a B grade. Site 4 at Burroo Bay and Site 6 at Griffins Bay were the only sites rated as being in fair condition. These condition results are very similar to the last reporting period. Importantly though trend analysis tells us what is happening on a long-term scale at these sites. Even though Burroo Bay (site 4) and Griffins Bay (site 6) were rated fair it is significant that Site 4 had decreasing trends for turbidity, total nitrogen and total phosphorous, despite this site being close to feeder creeks and catchment inputs. Site 6 had a decreasing trend for turbidity and no trend for nutrients and chlorophyll-a. Site 3 (near Picnic Island) is the only site in the Lake that continues to show an increasing trend for total nitrogen and total phosphorous. The in-lake sites continue to show no trend for chlorophyll a, total phosphorus, total nitrogen and turbidity values over the long term. This is not surprising though as they have been graded as being in good or very good condition since 2017.

Assessing recreational water quality (primary and secondary recreation contact) has been utilised through sampling for enterococci at three sites in Lake Illawarra since 2018. Estuarine sites are particularity impacted by potential sources of faecal contamination, including stormwater and urban and rural run-off. All sites had very similar percentage compliance to the previous year, which is a good result given the high rainfall that occurred over this summer compared to the lower rainfall over the previous 2022/23 summer period.

Monitoring the health of the lake should continue, as long-term datasets are essential to gain insights into how the lake is changing over time and is of greater value than focusing on specific individual sampling events. Given the large-scale developments occurring in the catchment of the lake and the potential impacts of a changing climate, it is important that the health of the lake is monitored. The implementation of the suite of management actions in the Lake Illawarra Coastal Management Program to control catchment inputs is strongly supported, and long-term monitoring of the lake will inform whether investment in the lake is making a difference.

Contents

Sum	nmary	1
Con	tents	2
1	Introduction	3
2	Water quality and estuary health	3
3	Monitoring program	4
	3.1 Water quality and estuary health monitoring	4
	3.2 Recreational water quality	5
4	Data Analysis	6
	4.1 Water quality analysis for estuary ecosystem health	6
	4.2 Estuary ecosystem health condition	7
	4.3 Assigning the Estuary Condition Grade	8
	4.4 Water quality for recreational use	9
5	Results	10
	5.1 Temporal analysis of parameters	10
	5.1.1. Temperature, salinity and pH	10
	5.1.2 Nitrite, Nitrate and Ammonia	11
	5.1.3 Total Nitrogen	11
	5.1.3 Phosphorus	15
	5.1.4 Chlorophyll-a	18
	5.1.5 Turbidity	19
	5.2 Water quality and estuary health trends	20
	5.3 Estuary ecosystem health condition	21
	5.4 Recreational water quality	24
6	Conclusion	25
Refe	erences	28
	endix 1: List of management actions in the Lake Illawarra Coastal Management Program relat	_
qqA	endix 2: Long-term plots of parameters at all sites from 2013/14 to April 2024	30

1 Introduction

Lake Illawarra is a significant natural asset for the community, and it is highly valued for its ecological, social and economic attributes (BMT 2020a). Wollongong City and Shellharbour City Councils, in partnership with the State Government, have prepared a Coastal Management Program to address the main threats to the Lake values. Catchment development and its potential impact on water quality in the Lake is identified to be one of the most significant threats needing to be managed (BMT 2020 b). Several actions in the Coastal Management Program for the Lake relate to protecting water quality. Targeted monitoring, evaluation and reporting of water quality and other health indicators are recommended to track the outcome of implementing these actions (refer to Appendix 1 for the list of water quality improvement and management actions in the Lake Illawarra Coastal Management Program 2020-2030).

There is a long history of water quality monitoring in the Lake, with various agencies involved at differing times. Wollongong City and Shellharbour City Councils took on this responsibility in October 2013, often with funding and technical assistance from the NSW Department of Climate Change, Energy, the Environment and Water (DCCEEW) coast and estuary management program, and since then regular reports on the results have been issued. These reports have considerable detailed analysis and show that there can be spatial and temporal variation in water quality, with season and weather patterns (particularly rainfall) having a major influence. This report covers the monitoring period from May 2023 to April 2024 and focusses on those measurements that are especially important for assessing the lake condition for its ecosystem health and recreational use.

This report has been developed to accompany the 2023-24 Lake Illawarra Estuary Health Report Card and provides the technical information on how the Report card scores are calculated as well as more detail on the results.

2 Water quality and estuary health

Good management of our estuaries requires an understanding of how estuaries function, an assessment of their condition and informed management decisions to maintain or improve the health of the estuary. The NSW Department of Climate Change, Energy, the Environment and Water (DCCEEW) (formerly Department of Planning and Environment) has undertaken extensive research of NSW estuary functioning, the impacts of catchment activities on estuary and the development of an estuary health monitoring and reporting framework (Roper et al 2011).

From this research the abundance of chlorophyll-a and turbidity are two measures recommended as indicators of estuary health. Wollongong and Shellharbour City Councils use these two indicators to measure the estuary health of Lake Illawarra, as well as a number of supporting water quality parameters such as nitrogen and phosphorous. Chlorophyll-a is an indicator of the microalgal abundance in a water body, and its measure is preferred for estuary health assessment as it is reported to be a good short-term indicator of response to a range of pressures and management, including nutrient (such as nitrogen and phosphorus) status. Water clarity and sediment inputs are assessed by measuring turbidity.

Ecological health does not refer to environmental health issues such as drinking water quality, safety for swimming and boating, heavy metal contamination or our ability to harvest shellfish or fish.

3 Monitoring program

3.1 Water quality and estuary health monitoring

Eleven sites within Lake Illawarra are monitored monthly for water quality and estuary health (Table 1, Figure 2). The parameters sampled are:

Water Quality:

- Temperature
- pH
- Dissolved oxygen
- Salinity
- Phosphorus (total phosphorous, filtered total phosphorous and filterable reactive phosphorus)
- Nitrogen (total nitrogen, nitrite and nitrate, and ammonia)

Estuary Health:

- Turbidity
- Chlorophyll-a



Algae

Algae or microscopic plants are always present in waterways but if conditions change and are suited to algal growth, blooms can occur. Blooms may occur if there are a lot of nutrients in the water which can come from urban stormwater, and fertiliser runoff from farms and gardens.

Algal blooms can reduce the amount of light reaching seagrass beds limiting their growth. When blooms of algae die and start to decay, the resulting bacterial activity can reduce oxygen concentrations in the water column, possibly leading to fish kills.

Chlorophyll-a

Chlorophyll-a is a pigment found in plants and is an essential molecule for the process of photosynthesis. In estuarine and marine waterways, chlorophyll-a is present in phytoplankton such as cyanobacteria, diatoms and dinoflagellates. As chlorophyll-a occurs in all phytoplankton it is commonly used as a measure of phytoplankton biomass (Roper et al 2011)



Sediment

Sediment from the land can be washed into waterways when it rains. Large amounts of sediment can come from roads and pathways washing directly into the stormwater, from poorly managed rural lands & unvegetated creek banks and foreshores. Too much sediment in the water reduces the amount of light reaching the bottom and is detrimental to seagrass which require light for growth.

Seagrass is critical for the health of estuaries as it provides essential habitat for fish and invertebrates. Excess amounts of suspended particles can also smother benthic organisms like sponges, irritate the gills of fish and transport contaminants.

Turbidity

Turbidity provides a measure of sediment in the water. It is the measure of light scattering by suspended particles in the water column, providing an indication of the amount of light penetration through the water column (Roper et al 2011).

Figure 1: Description of estuarine ecological indicators

Nitrogen is analysed as total nitrogen in unfiltered water (TN), the total after filtration (FTN), the amount present as nitrate and nitrite (often referred to as NOx's), and as ammonia, the reactive inorganic forms which are generally considered to be more bioavailable.

Phosphorus is analysed as total phosphorus in unfiltered water (TP), in filtered water (FTP), and as filterable reactive phosphorus (FRP). The filterable reactive phosphorus generally constitutes simple inorganic phosphorus (such as orthophosphate) and is considered more bioavailable than other forms of phosphorus.

The sampling procedure for turbidity and chlorophyll *a* are taken in accordance with the NSW Monitoring, Evaluation and Reporting sampling, data analysis and reporting protocols (State of NSW and Office of Environment and Heritage 2016).

3.2 Recreational water quality

Since the 2018/19 summer, three sites within Lake Illawarra were included for recreational water quality testing, following the NSW Beachwatch Program protocols, which test for the presence of enterococci – a group of bacteria indicating water quality condition for recreational use.

A site at the entrance (called Entrance lagoon beach) has been monitored for many years by the NSW Beachwatch Program and the data is analysed and reported within the NSW Beachwatch framework (State of NSW and DPE 2022). Currently the three sites added by Councils are not reported under the NSW Beachwatch program.

The sites added by Councils in 2018/19 to measure recreational water quality are located at Ski Way Park, Kanahooka and Purry Burry Point, which are popular launch sites for many recreational pursuits in the Lake (Table 1, Figure 2). These three sites were tested for the presence of enterococci on 19 occasions between October 2023 and April 2024.

Table 1: Description of the 11 sites monitored for water quality and estuary ecosystem health (in blue) and 3 sites monitored for recreational water quality (in orange)

Water Quality and Estuary Health sites			
Site ID	Site location	Lake Zone	
Site 2	Boat ramp at Windang Peninsula	Lake entrance	
Site 3	Bridge to Picnic Island	Lake entrance	
Site 3A	Jetty at Boonerah Point Reserve	Foreshore	
Site 4	Jetty at Sailing Club at Burroo Bay	Foreshore	
Site 5	Boat ramp and jetty at Kanahooka	Foreshore	
Site 6	Jetty at Griffins Bay Wharf	Foreshore	
NS1	North along a north-south transect	In-lake	
NS2	Middle along a north-south transect	In-lake	
NS3	South along a north-south transect	In-lake	
EW1	East along an east-west transect	In-lake	
EW2	West along an east-west transect	In-lake	
Recreational water quality sites			
Purry Burry Point	Primbee	Foreshore	
Ski Way Park	Oak Flats	Foreshore	
Kanahooka	Kanahooka/Koonawarra	Foreshore	

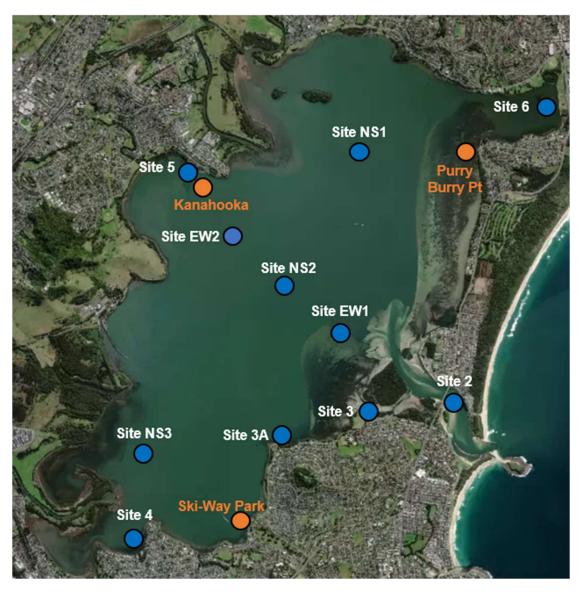


Figure 2: Map showing location of the 11 sites monitored by Council for water quality and estuary ecosystem health (in blue) and the 3 sites for recreational use (in orange)

4 Data Analysis

4.1 Water quality analysis for estuary ecosystem health

As in previous years, all indicators including those not covered in this report have been plotted against sampling date, rainfall, and the corresponding guideline trigger value from October 2013 to April 2024 for the 11 sites monitored. The indicators discussed in depth in this this report are various forms of nitrogen and phosphorous, chlorophyll-a and turbidity, which are some of the more important indicators of estuary ecosystem health and the catchment influence on the Lake. The guideline trigger values utilised are given in Table 2.

Table 2: Guideline trigger values

Parameter	Guideline	Source
Total Nitrogen (TN)	0.3 mg/L	ANZECC (2000)
Filtered Total Nitrogen (FTN)	0.3 mg/L	Based on TN from ANZECC (2000)
Nitrate and Nitrite (NOx's)	0.015mg/L	ANZECC (2000)
Ammonia	0.015 mg/L	ANZECC (2000)
Total Phosphorous (TP)	0.03 mg/L	ANZECC (2000)
Filtered Total Phosphorus	0.03 mg/L	Based on TP from ANZECC (2000)
Filtered Reactive Phosphorus	0.005 mg/L	ANZECC (2000)
Chlorophyll-a	3.6 μg/L ^a	State of NSW and DPE
Turbidity	5.7 NTU ^b	State of NSW and DPE

^a This value has been updated to 5 μg/L in State of NSW and Office of Environment and Heritage (2016)

The guideline trigger values for chlorophyll-a and turbidity continue to be the values previously adopted for the NSW Monitoring, Evaluation and Reporting Program (State of NSW and Office of Environment and Heritage 2013) rather than the updated values (State of NSW and Office of Environment and Heritage 2016), in order to maintain consistency with the values ulitised in earlier reports. These are also the values utilised in developing a risk-based framework for protecting the health of Lake Illawarra (Office of Environment and Heritage and the Environment Protection Authority 2017). Therefore, retaining these values as the desired target condition for the rest of the lake is reasonable at this time.

The data for TN, TP, chlorophyll-*a* and turbidity have also been subjected to a trend analysis using the water quality software program eWater to determine whether statistically significant trends are apparent for these indicators at any of the sites over the eight years. The non-parametric Seasonal Kendall test has been used for this, a method that is widely used to detect trends where there is a significant seasonal influence on water quality. Rainfall effects can detract from the seasonality pattern, and to account for this, data points that were greater than two standard deviations from the mean were excluded from the analysis. The trend analysis was performed with the filtered data.

4.2 Estuary ecosystem health condition

The estuary ecosystem health condition of each site has been determined based on its chlorophyll a and turbidity status over the summer months. The summer period is taken to be from 1 November to 30 April, while the winter is from 1 May to 31 October.

The methodology used is consistent with that recommended by the NSW Monitoring, Evaluation and Reporting (MER) protocols for estuaries and coastal lakes, which assesses the degree of compliance of these parameters with their water quality trigger values and allocates a condition grade ranging from very poor to very good, as described in Table 3. The full methodology for calculation of the estuary condition grade is described in State of NSW and Office of Environment and Heritage (2016). As noted in Table 2, the trigger values utilised for chlorophyll- α and turbidity are 3.6 μ g/L and 5.7 NTU respectively, rather than the updated values as reported in 2016 (State of NSW and Office of Environment and Heritage 2016).

^b This value has been updated to 6 NTU in State of NSW and Office of Environment and Heritage (2016)

4.3 Assigning the Estuary Condition Grade

The estuary condition grading system is based on detailed work by the NSW MER framework that looked at data for 130 estuaries in NSW. The grade definitions described in Table 3 are structured to allow easy comparison between different estuaries and for individual estuaries over time.

It is important that the cut-off values for each grade reflect the condition of each site in comparison to a broader scale of condition across all New South Wales estuaries (i.e., an 'Good' grade represents a good condition for a New South Wales estuary). To assist with the derivation of cut-offs, scores were calculated for 130 zones across a wide range of New South Wales estuaries using the same guidelines and worst expected values (State of NSW and Office of Environment and Heritage 2016). Cut-offs were then defined as representing a percentage of the scores for the state (Table 3). For example, a site score less than 0.07 defined the 20% of best site scores in the state and this became our 'Very Good' grade (see Figure 3 for other cut-offs).

Table 3: Descriptors for estuary ecosystem health condition grades

Grade	Result	Definition	Description
Α	Very Good	The indicator meets the benchmark	Equivalent to the best 20% of
		values for almost all of the time period.	scores in the state
В	Good	The indicator meets the benchmark	Equivalent to the next 30% of
		values for most of the time period.	good scores
С	Fair	The indicator meets the benchmark value	Equivalent to the middle 30% of
		for some of the time period.	scores
D	Poor	The indicator does not meet the	Equivalent to the next 15% of
		benchmark value for most of the time	poorer scores
		period.	
Е	Very Poor	The indicator does not meet the	Equivalent to the worst 5% of
		benchmark value for almost all of the	scores in the state
		time period.	

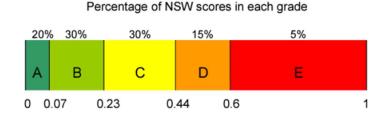


Figure 3: Relationship between distribution of NSW scores, grades and site scores

4.4 Water quality for recreational use

The analysis of the three sites within Lake Illawarra tested for the presence of enterococci calculates the percentage of the testing occasions when the sites complied with the guidelines for primary and secondary recreational use contact criteria (Table 4).

Table 4: Guideline trigger values for recreational use

Recreational Use	Guideline trigger value (enterococci)
Primary contact	35 cfu/100ml
Secondary contact	230 cfu/100ml

(source: ANZECC (2000))

5 Results

Several factors can influence water quality in a lake such as Lake Illawarra. These include weather, catchment runoff, assimilation and/or release of dissolved substances in the water by lake sediments, aquatic plants and animals, and the extent of flushing of the waterbody by tidal and catchment flows. These factors may not be uniform through the lake, suggesting variations in water quality can be expected in space and time. Rainfall has a significant influence on water quality and is considered one of the most important factors affecting water quality in the Lake in previous years. Table 5 presents the seasonal and yearly total rainfall records for the last twelve years and it shows that the 2023/24 period was significantly lower rainfall than the last two years, and similar to drier conditions experienced from 2018-2020. Compared to the year before, the largest amount of rainfall occurred during the summer period, with very low rainfall during the winter period (Table 5). It was the second lowest winter rainfall recorded since sampling began in 2009/10 (Table 2).

Table 5: Seasonal rainfall (mm) at Darkes Road MHL station since 2009

*Rainfall data from	Cleveland Road MHL station	

Year	Winter (1 May to 31 Oct)	Summer (1 Nov-30 April)	Year total
2009/10	333.5	523	856.5
2010/11	520	800	1320
2011/12	476.5	616	1092.5
2012/13	215	515	730
2013/14	498.5	813	1311.5
2014/15	365	771.5	1136.5
2015/16	461	460	921
2016/17	602.5	748	1350.5
2017/18	108	458	566
2018/19	253	407	660
2019/20	286.5	598	884.5
2020/21	474.5	720	1194.5
2021/22	384.5	1425.5	1810
2022/23	1046.5	510	1556.5
2023/24*	143	845	998

5.1 Temporal analysis of parameters

5.1.1. Temperature, salinity and pH

Long term graphs of temperature, salinity and pH since 2014 have been presented in Appendix 2 for the 11 sites monitored for water quality and estuary health. In the past 12 months sites have continued to show a seasonal pattern in temperature as evident in previous years (Appendix 2). The temperature variation between summer and winter can be as much as approximately 10-15°C, and this can be expected to cause seasonal change in other water quality processes which are temperature dependent.

A pH range of 7 to 8.5 is considered to be satisfactory for estuarine ecosystems (ANZECC 2000). Values did not go below or above this guideline values indicating, there are no concerns relating to pH at Lake Illawarra (Appendix 2)

Salinity since 2014 has been graphed against the daily rainfall records, as the salinity of the lake is influenced by freshwater flows, dilution and flushing rates of the estuary (Appendix 2). The results show that a salinity of around 30-35ppt continues to be maintained, except close to rainfall events where it decreases temporarily (Appendix 2).

5.1.2 Nitrite, Nitrate and Ammonia

When assessing the condition of an estuarine water body, chlorophyll α and turbidity are considered better indicators of estuary ecosystem health than the nitrogen and phosphorus concentrations. High nutrient concentrations do not always correlate with poor estuary health (Scanes et al. 2007), and there can often be a weak correlation between nitrogen loads and chlorophyll α concentrations (Roper et al. 2011). High nutrient inputs can, however, ultimately lead to poor water quality, and monitoring their concentrations in different parts of the lake can help identify inputs where nutrients may be significant and thus require management.

Over the May 2023 to April 2024 period the concentrations of nitrate and nitrite (known collectively as NOx) for the majority of sampling events for the in-lake sites were below or close to their respective detection limits of 0.01mg/L and below the guideline value of 0.015 mg/L) (Figure 4b). This indicates these more bio-available forms of nitrogen continue to be rapidly utilised by phytoplankton and other plant life in the Lake the majority of the time. Higher values were evident though related to two rainfall events that resulted in a significant nitrogen load to the lake; one at the beginning of December 2023 and one in April 2024. The foreshore sites showed greater variability and often higher values than the in-lake sites. High nitrate and nitrite levels were recorded at the foreshore sites during the same December and April rainfall events, but also during sampling at the end of September 2023. In particular, Site 3 and Site 6 displayed high levels of nitrate and nitrite for half of the sampling events (Figure 4a).

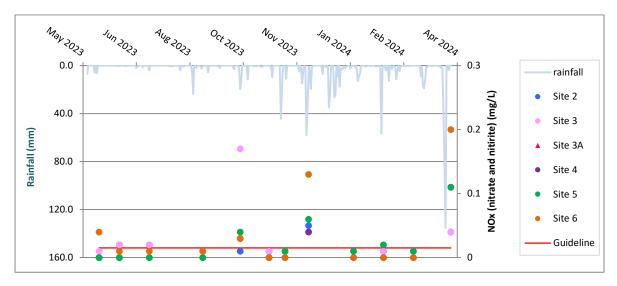
A similar pattern was noticeable for levels of ammonia (NH₃) in the lake (Figure 5). The in-lake samples were all below the recommended guideline value, except at NS1 during December 2023, that was related to a rainfall event (Figure 5a). The foreshore sites showed greater variability throughout the sampling period and much higher values of NH₃ levels, noticeably at Sites 3 and 6 indicating bioavailable nitrogen is occurring at these sites due to inputs from the catchment or foreshore, particularly during rainfall events (Figure 5b).

5.1.3 Total Nitrogen

Figure 6 shows total nitrogen (TN) values and rainfall for May 2023-April 2024 for the foreshore and in-lake sites. In Appendix 2 graphs of the results from the last ten years are presented for all 11 sites. Griffins Bay (Site 6) had TN values higher than the recommended guideline 10 out the 12 sampling events (Figure 6a). Site 4 also had high TN values for 75% of the sampling events, and this pattern of these two sites having high TN values is reflective of patterns in other years.

High total nitrogen values were related to rainfall and thus catchment inputs – exceedances above their recommended guideline occurred for the majority of foreshore sites in September and December 2023, and April 2024 (Figure 6a). The in-lake sites had values at the guideline values or exceed it during

A: Foreshore sites



B: In-lake sites

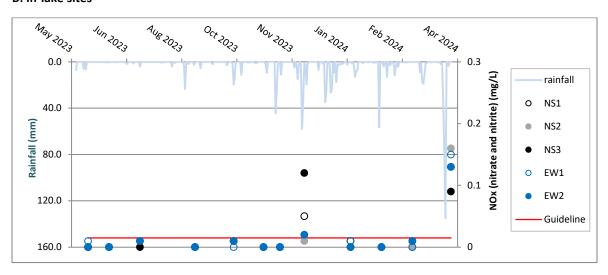


Figure 4: Plots of NOx (nitrate and nitrite) and rainfall from May 2023 to April 2024 for the (A) Foreshore (lake edge) sites and (B) In-lake sites

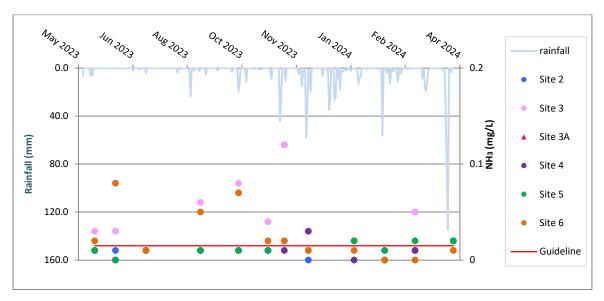
the December 2023 and Aril 202 rainfall events. All other sampling events though total nitrogen values were below the recommended guidelines value of 0.03mg/L (Figure 6b).

The TN value represents nitrogen that is present in water in both the dissolved and suspended forms, including microscopic algae and sediments, while the filtered total nitrogen (FTN) excludes the suspended component. The FTN for the May 2022-April 2023 period (Figure 7) continues to show better compliance with its guideline trigger value than TN. Similar to the pattern for TN, foreshore sites 3, 3A, 4, and 6 exceeded the ANZECC guideline during the rainfall event in September 2023, and

Site 6 also exceeded the guideline during the rainfall events in December 2023 and April 2024 (Figure 7a).

The in-lake sites showed compliance for the majority of the sampling period (Figure 7b). Only Site NS3 was just above the guideline value during the September rainfall event. Interestingly while most of the foreshore sites were below the guideline after the large April rainfall event, three in-lake sites exceed the guideline (Figure 7b).

A: Foreshore sites



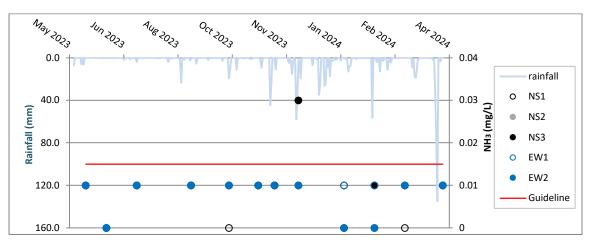
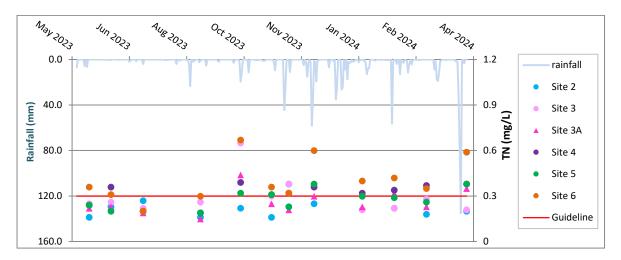


Figure 5: Plots of NH₃ and rainfall from May 2023 to April 2024 for the (A) Foreshore (lake edge) sites and (B) In-lake sites. Please note different NH₃ scales on each graph

A: Foreshore sites



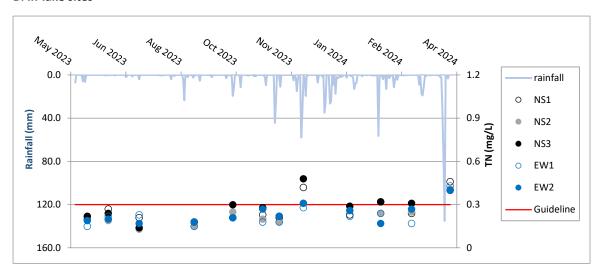
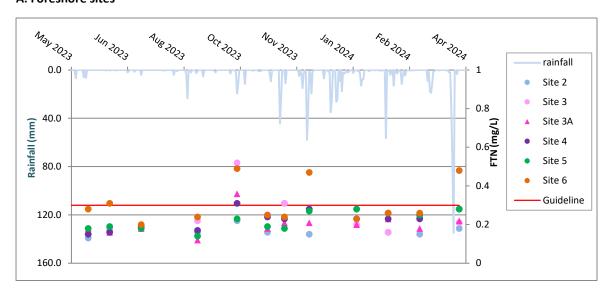


Figure 6: Plots of total nitrogen (TN) and rainfall from May 2023 to April 2024 for the (A) Foreshore (lake edge) sites and (B) In-lake sites

A: Foreshore sites



B: In-lake sites

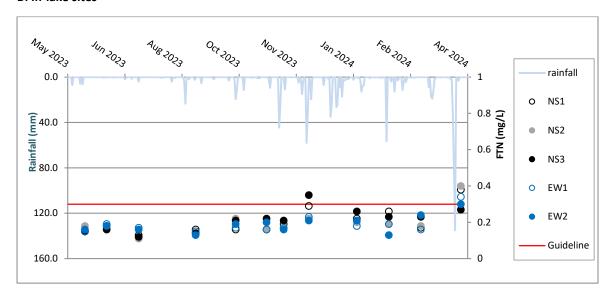


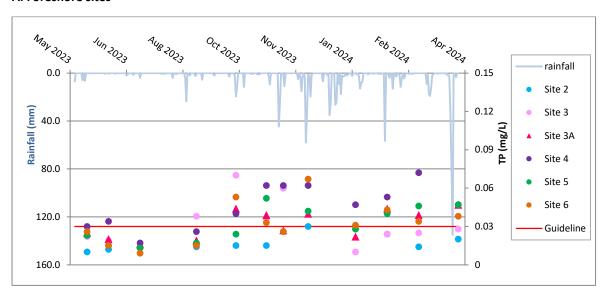
Figure 7: Plots of filtered total nitrogen (FTN) and rainfall from May 2022 to April 2023 for the (A) Foreshore (lake edge) sites and (B) In-lake sites

5.1.3 Phosphorus

Figure 8 shows total phosphorous (TP) values and rainfall for the period May 2023-April 2024 for the foreshore and in-lake sites. In Appendix 2 graphs of the results from the last nine years are presented. For the foreshore sites (Figure 8a) the recent observations are very similar to previous observations, with several sites routinely exceeding the guideline value, and with greater variation of total phosphorus values at the lake edge sites compared to in-lake sites. Along the lake edge Burroo Bay (Site 4) had high TP values for 8 of the 12 sampling events with very high values for all sampling events from October 2023 to April 2024 indicating a consistent source of phosphorus into this part of the lake. Sites 5 and 3A showed a similar pattern too (Figure 8a). Indeed, most foreshore sites (except

Sites 2 and 3) had high TP values in this date range as well, with readings 2-2.5x higher than the recommended guideline, largely a result of the high summer rainfall. Interestingly the large rainfall event on the 7th April 2024 did not see extremely high TP values when sampled five days later (Figure 8a). In-lake sites showed greater compliance with the guideline value, but most sites also showed higher readings from October 2023 to April 2024 due to rainfall (Figure 8b) and Site NS3 had a value 3x greater than the guideline in October 2023. The large rainfall event in April 2024 did see high TP values when sampled 5 days later but like the foreshore sites these were not extremely high values (Figure 8b).

A: Foreshore sites



B: In-lake sites

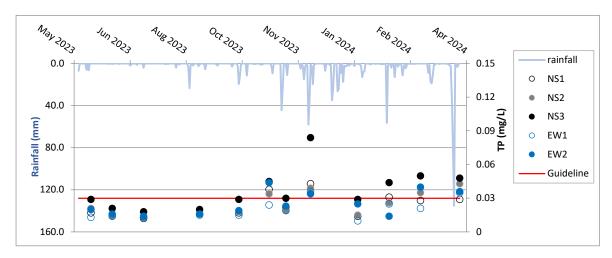
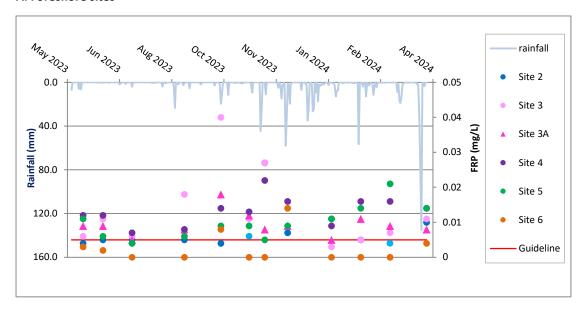


Figure 8: Plots of total phosphorous (TP) and rainfall from May 2022 to April 2023 for the (A) Foreshore (lake edge) sites and (B) In-lake sites

Lake Illawarra is known to be a phosphorus-rich environment due to the catchment soils having naturally high phosphate levels, which may explain the high values in the absence of very high rainfall events.

The filterable reactive phosphorus (FRP) results are presented in Figure 9. The guideline value for this form of phosphorus is very low at 0.005 mg/L, and in a phosphorus-rich environment such as Lake Illawarra most sites continue to exceed it. Similar to total phosphorous most foreshore sites showed higher values from October 2023 to April 2024, and greatly exceeded the guideline (Figure 9a). Interestingly Site 6, which is commonly in fair condition had some of the lowest values and was well below the guideline for most sampling events (Figure 9a). Most of the in-lake sites were below or slightly exceeded the guideline value (Figure 9b). Site NS3 showed the greatest exceedances throughout the sampling period. At most sites, about 70 to 80% of the total phosphorus (TP) in the water is present in the dissolved form (FTP), and about half of this dissolved fraction is in the reactive form (FRP). The detection of this reactive form of phosphorus in the water suggests that phosphorus is not a limiting nutrient for primary production in the lake.

A: Foreshore sites



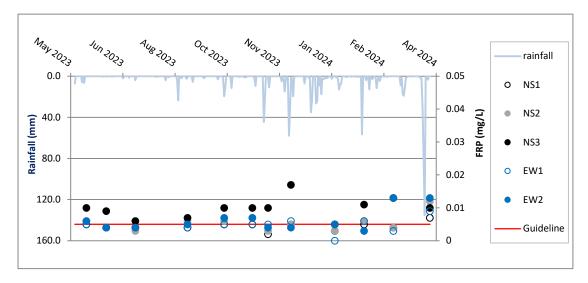


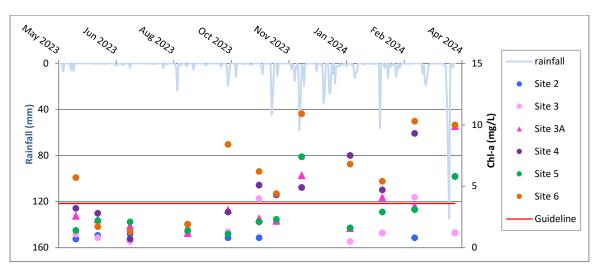
Figure 9: Plots of filterable reactive phosphorus (FRP) and rainfall from May 2023 to April 2024 for the (A) Foreshore (lake edge) sites and (B) In-lake sites

5.1.4 Chlorophyll-a

Between May to September 2023 all foreshore sites, except Site 6 on one occasion, were below the recommended guideline value (Figure 10a). From October 2023 to April 2024 though there was a much higher values and variability in chlorophyll-a with foreshore site regularly exceeding the guideline, where spring and summer rainfall was much higher than the autumn and winter months. This was particularly evident for Sites 4 and 6 (Figure 10b).

In-lake sites showed greater compliance with the NSW MER trigger value for the majority of the sampling events (Figure 10b). Exceedances were recorded during the December 2023 and April 2024 sampling events, which coincided with high rainfall events (Figure 10b).

A: Foreshore sites



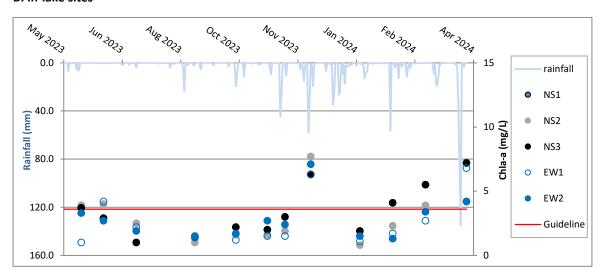
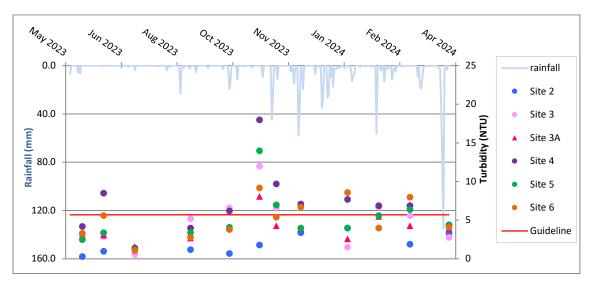


Figure 10: Plots of chlorophyll-a and rainfall from May 2023 to April 2024 for the (A) Foreshore (lake edge) sites and (B) In-lake sites

5.1.5 Turbidity

Turbidity gives an indication of light availability in the water and is influenced by the amount of suspended material present. In a relatively shallow lagoon such as Lake Illawarra, together with the amount of suspended microscopic algae and other organisms, the nature of the bottom sediments (whether muddy or sandy), the weather conditions (especially wind and rain), and boating or other traffic that can cause local turbulence in the water are all important factors that can affect the turbidity levels, as well as catchment run-off.

A: Foreshore sites



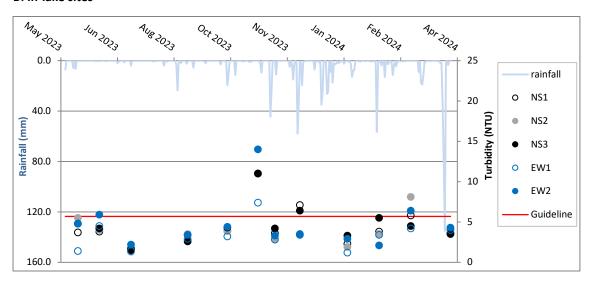


Figure 11: Plots of turbidity and rainfall from May 2023 to April 2024 for the (A) Foreshore (lake edge) sites and (B) In-lake sites

The results for the last 12 months show for the majority of the sampling period foreshore sites and in-lake sites were below or just above the trigger value, except for the October 2023 sampling event. Very high values were recorded at all sites (except Site 2 at the entrance) during the October 2024 sampling event, and interestingly it did not appear to be related to rainfall (Figure 11a). While there was greater variability in turbidity values at the foreshore sites, the in-lake sites were generally just above or below the guideline value, except for the October 2023 sampling event (Figure 11b). The foreshore sites are shallower than the in-lake zones, and are dominated by muddy bottom sediments, except around the eastern margin where sediments are sandier. These mud-dominated edge sites can generally be expected to be more turbid than the sites around the entrance and in the deeper in-lake zones, as any turbulence in the water caused by wind or boating activities can quickly mobilise the bottom sediments. Also, the foreshore sites are more impacted by catchment run-off.

Interestingly the large rainfall event in April 2024 did not see exceedance in turbidity recorded at any site (Figure 11).

5.2 Water quality and estuary health trends

Water quality trends over time are important because they can inform whether management strategies put in place to protect the health of the lake are effective. Data over a reasonably long period is required for these trends to become apparent, as there can be significant short-term variation arising from seasonal and meteorological effects, and these can detract from the background trend. Whether these factors are significant must be considered in any trend analysis.

The trend analysis was performed with the filtered data for total nitrogen, total phosphorous, turbidity and chlorophyll-a. A decreasing trend means that the values are decreasing over time and hence an improvement in the condition of that site.

Foreshore sites are particularly vulnerable to catchment run-off, and it is significant that Sites 3A, 4, 5 and 6 have decreasing trends for at least one parameters despite them being close to feeder creeks and catchment inputs (Table 6). In particular Site 4 has decreasing trends for turbidity, total nitrogen and total phosphorous. Site 3 is the only site in the Lake that continues to show an increasing trend for total nitrogen and total phosphorous, which is the same result as last year's sampling period. The reason for this could be that increased sand deposition across the seaward end of this channel observed over time is restricting flushing of this area, plus the presence of a large stormwater drain in the vicinity.

The in-lake sites continue to show no trend for chlorophyll *a*, total phosphorus, total nitrogen and turbidity values over the long term (Table 6). This is not surprising though as they have been graded as being in good or very good condition since 2017 (refer to Figure 12).

Table 6: Results of trend analysis for chlorophyll-a, turbidity, total nitrogen and total phosphorous at all sites

Site	Chlorophyll a	Turbidity	Total nitrogen	Total phosphorous
2	No trend	No trend	No trend	No trend
3	No trend	No trend	Increasing trend	Increasing trend
3A	No trend	Decreasing trend	Decreasing trend	No trend
4	No trend	Decreasing trend	Decreasing trend	Decreasing trend
5	No trend	Decreasing trend	No trend	No trend
6	No trend	Decreasing trend	No trend	No trend
NS1	No trend	No trend	No trend	No trend
NS2	No trend	No trend	No trend	No trend
NS3	No trend	No trend	No trend	No trend
EW1	No trend	No trend	No trend	No trend
EW2	No trend	No trend	No trend	No trend

5.3 Estuary ecosystem health condition

The estuary ecosystem health condition is based on the chlorophyll a and turbidity data for the summer period only (November to March), as recommended under the NSW MER framework for estuaries and coastal lakes (State of NSW and Office of Environment and Heritage 2016), using the guideline trigger values of 3.6 μ g/L for chlorophyll-a and 5.7 NTU for turbidity. The results for the recent summer (2023/24) are presented in Figure 12. Results from the previous years from each summer period from 2015/16 are shown in Figure 13 so changes in condition over the years can be seen. There have been changes in the number and location of some sites over the years.

The results show that for the 2023/24 period, four of the eleven sites are in very good condition and five sites in good condition (Figure 12). Two foreshore sites – Site 6 in Griffins Bay and Site 4 at Burroo Bay were rated as being in fair condition. An overall estuary health condition grade has been calculated for Lake Illawarra. The estuary was rated as being in good condition or a B grade (Figure 12), and this is the same result as last year.

Site 6 has continuously been rated as fair to very poor since 2013/14, indicating the high input of nutrients and sediment from the catchment in this area and the lack of flushing that occurs in this bay.

In the previous sampling period Site 4 (Burroo Bay) was graded was also graded as fair (Figure 13). This was largely due to several high turbidity values that were double the recommended guideline value. This site is situated near to two catchment inputs – Macquarie Rivulet and Horsley Inlet which could be the source of the high turbidity values. Also, it is located at the Oak Flats Sailing Club and turbulence from activities at the foreshore could be contributing to high turbidity values.

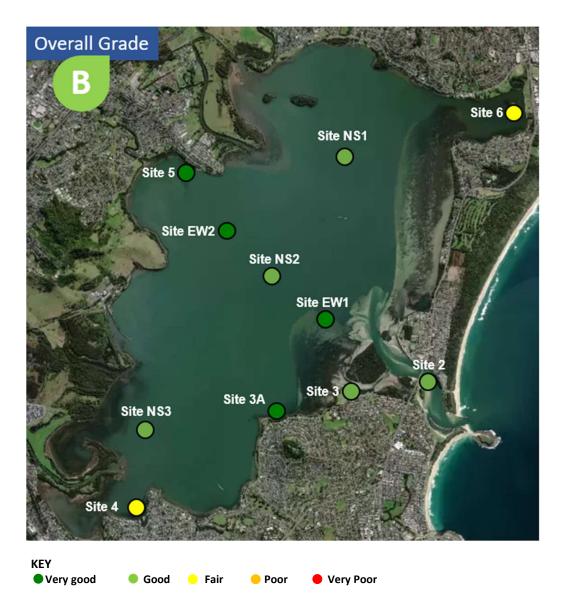


Figure 12: Estuary health condition ratings (based on chlorophyll *a* and turbidity) over the summer period (November to April) for 2023/24 for 11 sites monitored across Lake Illawarra, and the overall estuary grade for Lake Illawarra.

Site 5 last year was rated as being in fair condition largely a result of several very high chlorophyll *a* and turbidity values over the summer period related to rainfall events. This year it was rated as very good condition. This is a good result as there was more rainfall this summer than the last summer period on which estuary condition grades are based, and hence more catchment runoff from Brooks Creek that is located near this site.

In past years other foreshore sites have had poorer estuary health condition ratings compared to the in-lake sites and was related to these lake edge sites being more susceptible to catchment inputs (Figure 13). Turbidity at edge sites can also be easily influenced by prevailing wind conditions at the time of sampling, particularly as the lake is very shallow, and by other disturbances such as boating. However, in the last few years there has been an improvement in the health of foreshore sites, with the majority of the foreshore sites now being rated as being in good condition (Figure 12). Plots of the estuary health condition of the sites since 2016 show that over time the condition of the foreshore sites within Lake Illawarra has improved (Figure 13).

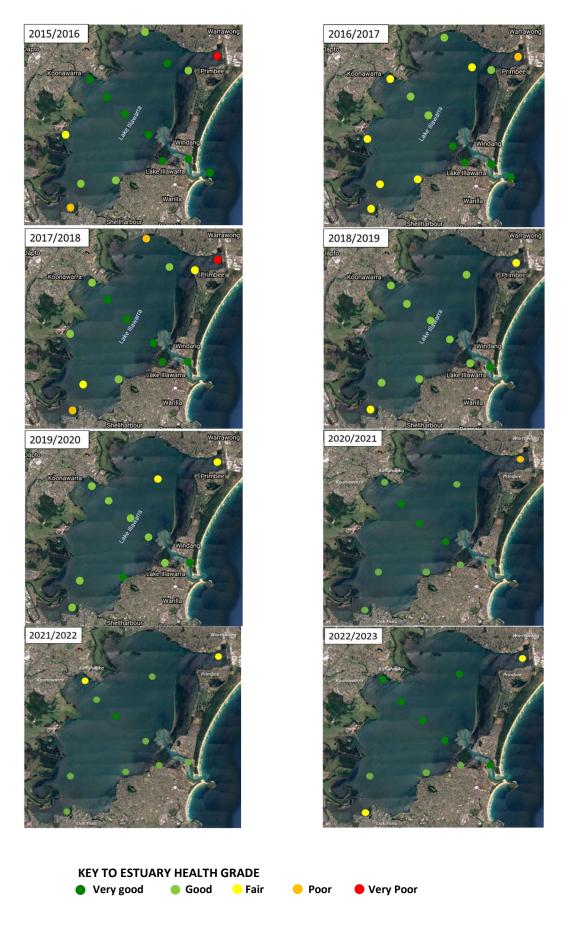


Figure 13: Estuary health condition ratings (based on chlorophyll a and turbidity) over the summer period (November to April) for each year from 2015-2023 for 11 sites monitored across Lake Illawarra.

5.4 Recreational water quality

Figure 14 shows the percentage compliance of the 20 test occasions meeting the recreational water guideline for primary (35 cfu/100 ml) and secondary (230 cfu/100 ml) contact for three estuarine sites sampled for the presence of enterococci. Estuarine sites are particularity impacted by potential sources of faecal contamination, including stormwater, urban and rural run-off. All sites had a higher percentage compliance than the previous year which is surprising given the amount of summer rainfall (Figure 14 and 15). Secondary recreational contact ranged from 89-95%, whereas primary recreational contact compliance ranged from 74% at Ski-Way Park to 84% and 89% at Kanahooka and Purry Burry Point respectively (Figure 14).

Rainfall is the major driver of pollution and enterococci levels to recreational water quality, and it is recommended that swimming should be avoided during and for up to one day following heavy rain at ocean beaches and up to three days at estuarine sites (NSW DPIE 2022). High values were associated with rainfall, especially during December sampling event, and a high rainfall event in February saw high values at Ski-wat Park and Purry Burry (Figure 14). Sites were sampled 12 days after the very high rainfall event in April 2024, and sites were below secondary recreation levels, and except for Purry Burry, under the primary recreation guideline too, suggesting any runoff had been flushed from the estuary by then.

The entrance site (called Entrance Lagoon Beach) that is sampled, analysed and reported under the NSW Beachwatch program was rated as good for the 2022-2023 period (State of NSW and DPE 2022). The beach suitability grade of good means that the location has generally good microbial water quality and water is considered suitable for swimming most of the time (State of NSW and DPE 2022).

No NSW Beachwatch summary report was available for the 2023-2024 summer period at the time of publication of this report.

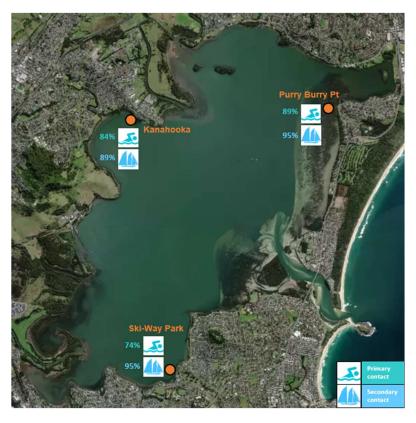


Figure 14: Percentage of the test occasions meeting primary (35 cfu/100 ml) and secondary (230 cfu/100ml) contact recreational water quality guidelines, at three locations within Lake Illawarra from October 2023-April 2024.

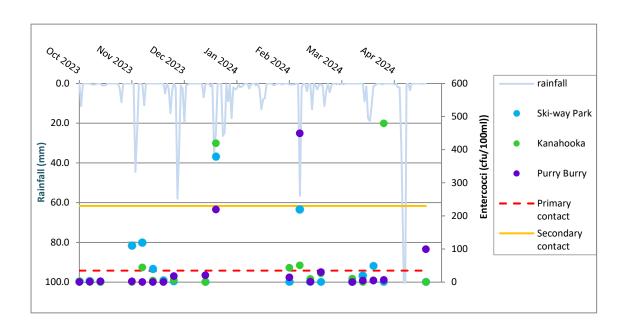


Figure 15: Plot of enterococci concentrations against rainfall from October 2023-April 2024

6 Conclusion

This report has reviewed selected parameters and indicators to describe the water quality and estuary health condition in Lake Illawarra for the period May 2023 to April 2024, and for trends in estuary health since 2013. Recreational water quality has also been monitored over the summer period. The indicators are nitrogen and phosphorus concentrations, turbidity and chlorophyll- α for estuary health, and the bacterial species enterococci for recreational water quality.

A diverse range of factors can influence water quality including residence times as set by mainly circulation patterns and freshwater inputs, sediment fluxes and geomorphological processes. Evidence for strong spatial variation in nutrient values and turbidity around the lake, as noted in previous reports, has been reinforced with recent data. Water quality at the foreshore edge sites show greater variation and more exceedances of the ANZECC guidelines or trigger values than the in-lake sites. This is to be expected given the lake edge sites are closer to the freshwater catchment inputs and are not as greatly flushed as the in-lake sites. Also, the foreshore sites are generally shallower than the in-lake zones, and are dominated by muddy bottom sediments, except around the eastern margin where sediments are sandier. These mud-dominated edge sites can generally be expected to be more turbid than the sites around the entrance and in the deeper in-lake zones, as any turbulence in the water caused by wind or boating activities can quickly mobilise the bottom sediments. The spatial differences also reinforce that the residence time for the more central portion of Lake Illawarra is shorter than that for the more enclosed areas of the lake (e.g., Griffins Bay). Longer residence times in the more enclosed areas north in the lake and some sections in the south, where water quality can be poor, mean that catchment inputs have more time to be incorporated into bottom sediments. These sediments are part of the internal nutrient reserves that feed the water column, contributing in part to the poor water quality observed in these areas, particularly at Site 6 Griffins Bay and Site 4 Burroo Bay.

Temporal variations in water quality can be evident over different timescales. In past years this seasonal differences have not been very apparent, largely because there was high rainfall during the

winter months. The results for this sampling period show seasonal differences were apparent. Over the summer months (November to April) when temperatures and daylight hours are greater than in the winter months (May to November), higher nutrient and chlorophyll-a concentrations are known to be present in estuarine waters. This seasonal pattern is not uncommon and has been observed in other waterbodies, regardless of rainfall conditions. This was noticeable for nutrients and chlorophylla during this sampling period but was also related to the much higher rainfall over the summer compared to the winter months. Rainfall and hence freshwater discharge continues to be an influential factor on the water quality of the Lake.

Based on current analysis the condition of the estuary seems to be stable with similar results to last year with nine of the eleven sites rated good to very good for 2023/24, albeit with a few less sites being in very good condition. This can be attributed to the rainfall over the summer increasing the chlorophyll-a and turbidity values over the summer. The Lake was graded as being in Good condition or a B rating, which was the same grading as last year (Wollongong City Council 2023). Lake Illawarra appears resilient and maintaining good estuary health. It is important to note that this is based on only two indicators of estuary health and other ecological impacts may be occurring.

Two foreshore sites were rated as Fair—Site 6 in Griffins Bay and Site 4 at Burroo Bay. Site 6 has continuously been rated as fair to very poor since 2013/14, indicating the high input of nutrients and sediment from the catchment in this area and the lack of flushing that occurs in this bay. Griffins Bay (Site 6) had TN values higher than the recommended guideline 10 out the 12 sampling events and Site 4 also had high TN values for 75% of the sampling events. This pattern of these two sites having high TN values is reflective of patterns in other years. Site 6 also had several high chlorophyll-a values over the summer period that were double the recommended guideline value. Site 4 was also rated fair largely due to several high turbidity values that were double the recommended guideline value. This result could be due to a number of influences. It is situated near to two catchment inputs – Macquarie Rivulet and Horsley Inlet which could be the source of the high turbidity values. Also, it is located at the Oak Flats Sailing Club and turbulence from activities at the foreshore would be contributing to high turbidity values. Lastly it is in a slightly enclosed bay and hence would not receive much flushing and the residence time would be greater.

Importantly though the trend analysis tells us what is happening on a long-term scale at these sites. Trend analysis was performed for total nitrogen, total phosphorous, turbidity and chlorophyll-a. A decreasing trend means that the values are decreasing over time and hence an improvement in the condition of that site. Water quality trends over time are important because they can inform whether management strategies put in place to protect the health of the lake are effective. Data over a reasonably long period is required for these trends to become apparent, as there can be significant short-term variation arising from seasonal and meteorological effects, and these can detract from the background trend.

Even though Sites 4 and 6 were rated fair it is significant that these sites still show a decreasing trend for some parameters, despite them being close to feeder creeks and catchment inputs. Site 4 having a decreasing trend for total nitrogen, total phosphorus and turbidity. Site 6 and Site 5 have a decreasing trend for turbidity, and Site 3A has a decreasing trend for turbidity and total nitrogen. Site 3 is the only site in the Lake that continues to show an increasing trend for total nitrogen and total phosphorous. The reason for this could be that increased sand deposition across the seaward end of this channel observed over time is restricting flushing of this area, plus the presence of a large stormwater drain in the vicinity.

The in-lake sites continue to show no trend for chlorophyll-*a*, total phosphorus, total nitrogen and turbidity values over the long term. This is not surprising though as they have been graded as being in good or very good condition since 2017.

Recreational water quality was monitored over the summer period at three sites within the lake and the percentage compliance with primary and secondary recreation contact compliance was calculated. Estuarine sites are particularity impacted by potential sources of faecal contamination from animal faeces and sewerage overflows, including urban and rural stormwater run-off. All of the three sites had higher percentage compliance than the previous year (WCC 2023). This is a surprising result given rainfall was quite high over the summer period. Rainfall is the major driver of pollution and high enterococci levels to recreational water quality, and it is recommended that swimming should be avoided up to three days at estuarine sites following heavy rain (State of NSW and DPE 2022).

Continuation of the monitoring program will help establish how the lake responds following extended wet weather periods and the impact of catchment run-off, as well as gaining insights to how the estuary is varying over time and responding to changing climate regimes. Climate change is expected to cause heavy rainfall events to become more extreme with an increase in east coast lows over the warmer months. This will lead to increased catchment run-off, periods of decreased salinity in estuarine waters, and increased sediment and nutrient loads in coastal and estuarine waters (Adapt NSW 2022).

As part of the implementation of the Lake Illawarra Coastal Management Program, Wollongong and Shellharbour City Councils implementing a catchment water quality and ecological health program. This program captures baseline water quality monitoring data for the catchment, including from wet weather events for a number of environmentally and aesthetically important urban and rural creeks within the Lake Illawarra catchment. The program is also sampling macroinvertebrates and benthic diatoms to give an indication of the ecological health of these creeks, This data will inform us in understanding inputs into the lake and in current and future environmental management and monitoring programs.

References

Adapt NSW (www.climatechange.environemnt.nsw.gov.au)

ANZECC (2000) "Australia and New Zealand Guidelines for Fresh and Marine Water Quality."

BMT (2020a) "Community Uses, Values, Threats and Opportunities Lake Illawarra."

BMT (2020b) "Lake Illawarra Coastal Management Program – 2020-2030."

Glamore, W. C., D. S. Rayner, and P. F. Rahman (2016) Estuaries and climate change. Technical Monograph prepared for the National Climate Change Adaptation Research Facility. Water Research Laboratory of the School of Civil and Environmental Engineering, UNSW

Office of Environment and Heritage and the Environment Protection Authority (2017) "Risk-based framework for considering waterway health outcomes in strategic landuse planning decisions."

Roper T, Creese B, Scanes P, Stephens K, Williams R, Dela-Cruz J, Coade G, Coates B & Fraser M. (2011) Assessing the condition of estuaries and coastal lake ecosystems in NSW, Monitoring, evaluation and reporting program, Technical report series, Office of Environment and Heritage, Sydney

Scanes P, Coade, G., Doherty, M., & Hill, R. (2007). Evaluation of the utility of water quality-based indicators of estuarine lagoon condition in NSW, Australia. Estuarine, Coastal and Shelf Science, 306-19.

State of NSW and Office of Environment and Heritage (2013) "Assessing Estuary Ecosystem Health: Sampling, data analysis and reporting protocols."

State of NSW and Office of Environment and Heritage (2016) "Assessing estuary ecosystem health: sampling, data analysis and reporting protocols."

State of NSW and Department of Planning and Environment (DPE) (2022) Beachwatch State of the Beaches 2021-2022.

Wollongong City Council (2018) "Lake Illawarra Water Quality Monitoring and Reporting."

Wollongong City Council (2021) "Lake Illawarra Estuary Health and Water Quality Report."

Wollongong City Council (2022) "Lake Illawarra Estuary Health and Water Quality Report."

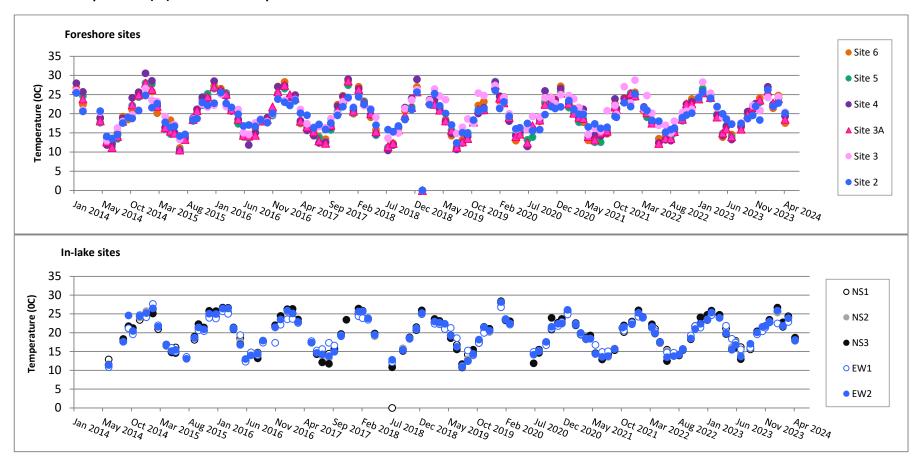
Wollongong City Council (2023) "Lake Illawarra Estuary Health and Water Quality Report."

Appendix 1: List of management actions in the Lake Illawarra Coastal Management Program relating to improving water quality

Strateg	Strategy 1: Improve Water Quality (WQ)				
WQ1	Implement a Risk Based Stormwater Management Framework for the Lake Illawarra catchment				
WQ2	Upgrade existing stormwater quality management measures, or install new devices, which may				
	include water sensitive urban design or other design that will improve water quality as well as				
	entrance habitat and natural values				
WQ3	Review and prioritise maintenance and cleaning regime for existing stormwater quality devices				
WQ4	Design and implement targeted catchment input monitoring as required for developments				
	resulting in a large-scale change or intensification of land use				
WQ5	Reduce sediment load to the Lake by improving compliance with erosion & sediment controls				
	for development sites				
WQ6	Reduce the impact of sewer overflows				
WQ7	Implement water quality monitoring programs for estuary health, recreational use and physico-				
	chemical and bacteriological indicators in the Lake and its catchment				
WQ8	Improve litter management				
WQ9	Investigate and manage potential pollution sources including contaminated sites that				
	contribute to poor water quality in the Lake.				

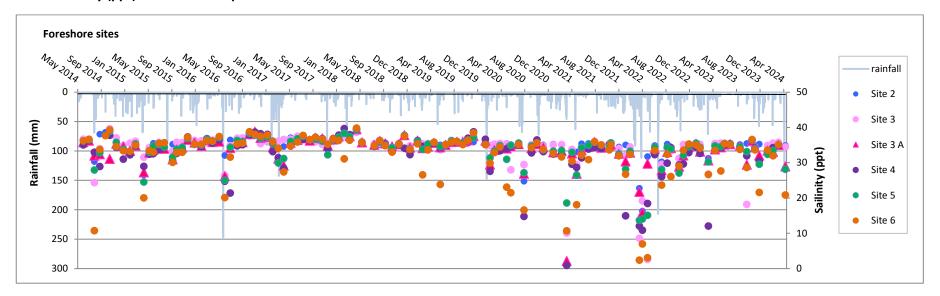
Appendix 2: Long-term plots of parameters at all sites from 2013/14 to April 2024

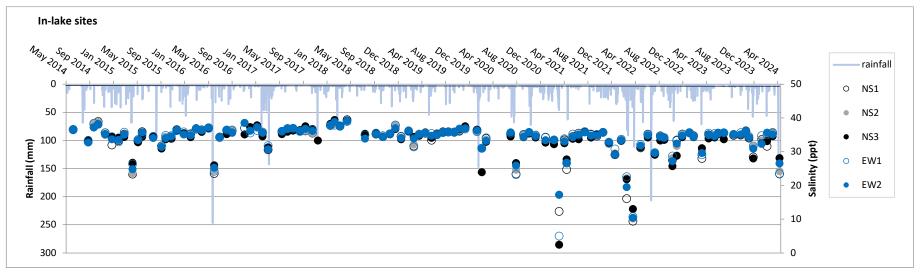
Plots of Temperature (°C) from 2014 to April 2024



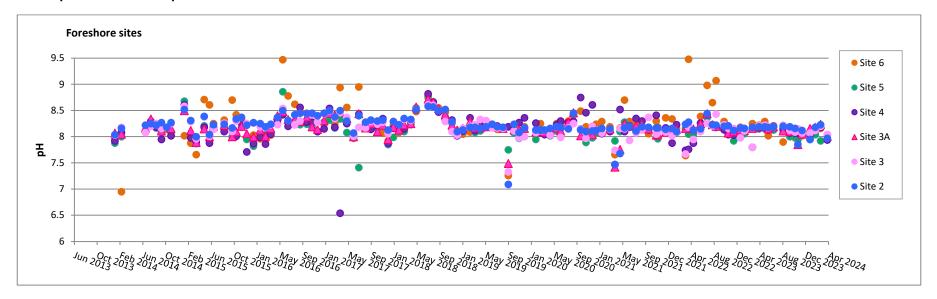
Note: zero readings due to equipment issues

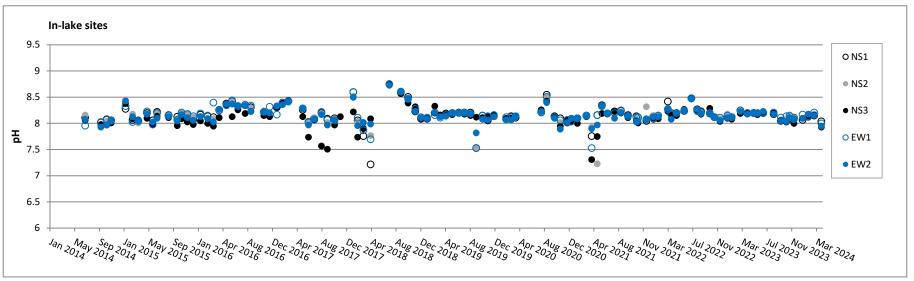
Plots of Salinity (ppt) from 2014 to April 2024



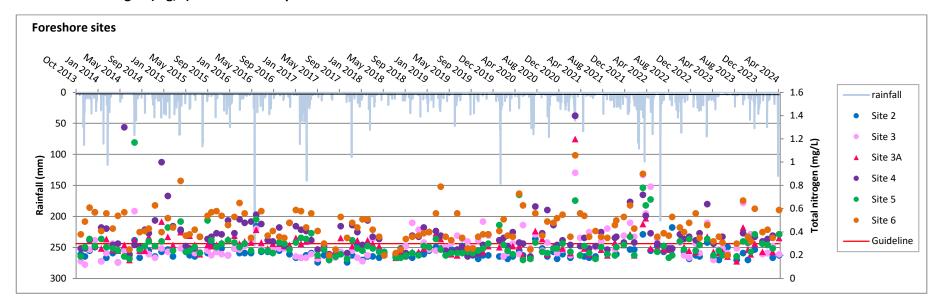


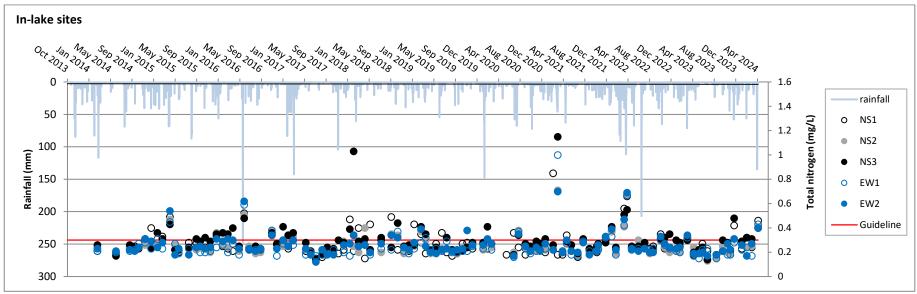
Plots of pH from 2014 to April 2024



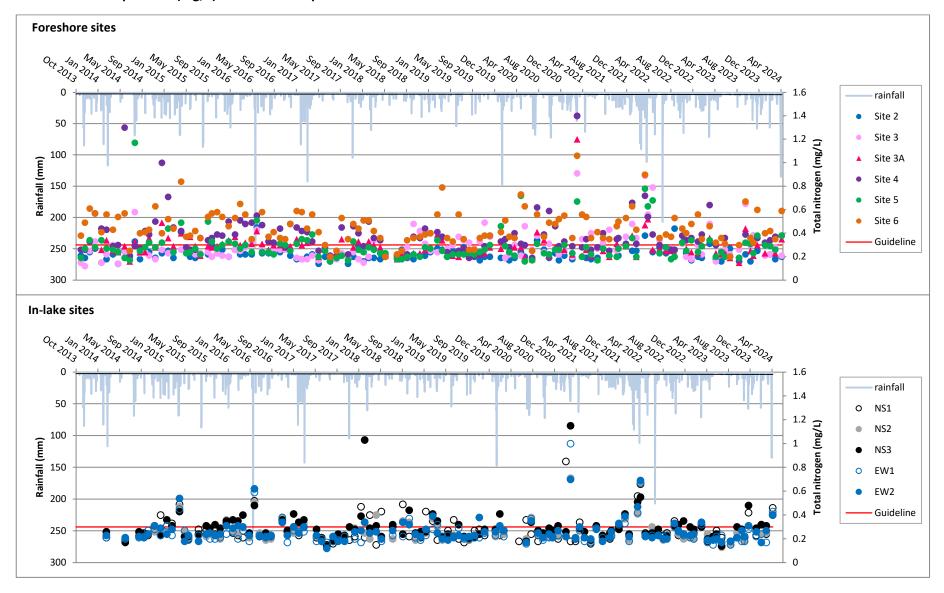


Plots of Total Nitrogen (mg/L) from 2013 to April 2024

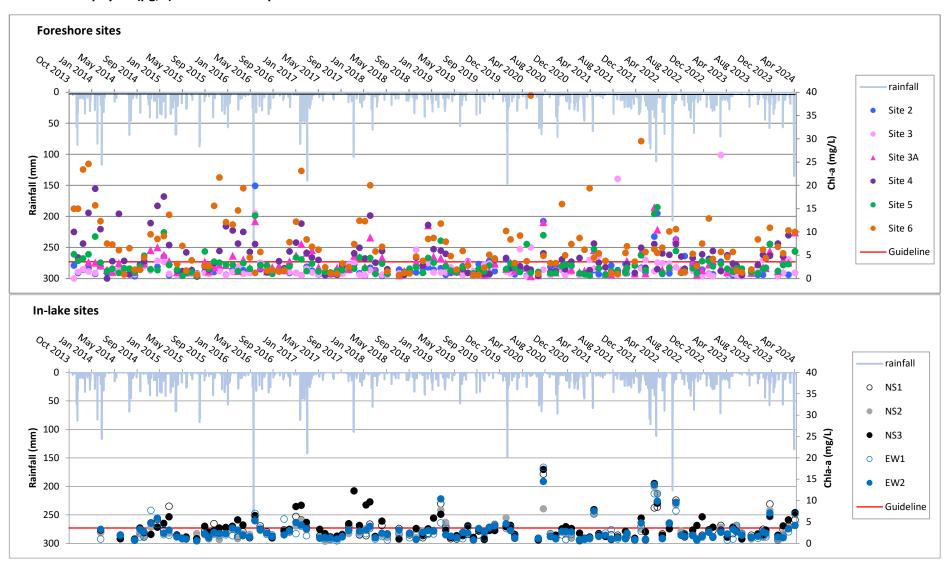




Plots of Total Phosphorous (mg/L) from 2013 to April 2024



Plots of Chlorophyll a (µg/L) from 2013 to April 2024



Plots of Turbidity (NTU) from 2013 to April 2024

